## California Department of Fish and Wildlife North Central Region

Middle Sacramento River Juvenile Salmon and Steelhead Monitoring Project

## Timing, Composition, and Abundance of Juvenile Salmonid Emigration in the Sacramento River Near Knights Landing September 2015 - June $2016{ }^{1}$



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## LIST OF ABBREVIATIONS AND ACRONYMS

| BBY | Bismark Brown |
| :--- | :--- |
| BO | Biological Opinion |
| BY | Brood year |
| CA | California |
| CAMP | Comprehensive Assessment and Monitoring Program |
| CDEC | California Data Exchange Center |
| cfs | Cubic feet per second |
| CDFW | California Department of Fish and Wildlife |
| cm | Centimeter |
| CNFH | Coleman National Fish Hatchery |
| CPUE | Catch per unit of effort |
| CVP | Central Valley Project |
| CWT | Coded wire tag |
| DCC | Delta Cross Channel |
| FL | Fork length |
| km | kilometer |
| LAD | Length at date |
| LSNFH | Livingston Stone National Fish Hatchery |
| mm | Millimeter |
| NMFS | National Marine Fisheries Service |
| NTU | Nephelometric turbidity units |
| QAQC | Quality assurance and quality control |
| rkm | River kilometer |
| RST | Rotary screw trap |
| SD | Standard deviation |
| SWP | State Water Project |
| USFWS | United States Fish and Wildlife Service |
| VIE | Visible Implant Elastomer |
| YOY | Young-of-the-year |

## EXECUTIVE SUMMARY

The North Central Region of the California Department of Fish and Wildlife operates a juvenile salmonid monitoring program on the Sacramento River in California (CA) to obtain information on the temporal distribution, relative abundance, and run composition of juvenile Chinook salmon (Oncorhynchus tshawytscha) and steelhead trout (O. mykiss) emigrating from the upper Sacramento River to the Sacramento-San Joaquin Delta (Delta). These data are collected at two separate locations that use two paired rotary screw traps (RST) outfitted with 28 -foot cones. The most downstream location is 0.8 kilometers (km) downstream of Knights Landing, CA, at Sacramento River kilometer (rkm) 144. Data collection is permitted under an Endangered Species Act Section 10(a)(1)(A) Permit (14808-4M) issued by the National Marine Fisheries Service (NMFS).

The monitoring program entered its $20^{\text {th }}$ consecutive year of sampling at the Knights Landing monitoring site beginning on September 16, 2015. Sampling concluded on June 20, 2016, for a total of 280 days of sampling.

During the season, 17,745 unmarked (adipose fin intact) juvenile Chinook salmon were captured. Peak catch occurred during calendar week 2 through 4, when 13,948 unmarked juvenile Chinook were captured. Juvenile Chinook salmon were identified to run using length-at-date (LAD) criteria developed by Fisher (1992) and modified by Greene (1992). The LAD run assignment is a widely used technique in the Central Valley for identifying juvenile Chinook salmon when multiple runs are present (Harvey 2011). Of the 7,912 unmarked juvenile Chinook salmon captured, 51 ( $0.2 \%$ ) were assigned to winter-run, 926 ( $5.2 \%$ ) were assigned to springrun, $16,767(94.5 \%)$ were assigned to fall-run, and $1(<0.1 \%)$ were assigned to late fall-run. Trap efficiency data were applied to catch totals to produce run-specific passage estimates. The passage estimate for fall-run was $8,434,470$; for spring-run was 357,030 ; and for winter-run was 11,472 . An estimate was not produced for late fall-run Chinook because too few individuals were captured to create a reliable estimate.

A total of 469 hatchery origin Chinook salmon was captured by the Knights Landing RSTs. These fish were identified by a missing adipose fin which is removed by hatchery staff prior to fish release. During the sampling period, releases of brood year (BY) 2015 fall- and late fall-run Chinook salmon were completed by Coleman National Fish Hatchery (CNFH). Additionally, releases of BY 2015 winter-run Chinook salmon were completed by Livingston Stone National Fish Hatchery (LSNFH). These releases occurred upstream of the Knights Landing sampling site. Of the 469 hatchery origin Chinook salmon captured, 52 (10.7\%) were identified as winter-run, 55 (11.7\%) were identified as late fall-run, 159 (33.9\%) were identified as spring-run, and 203 ( $42.3 \%$ ) were identified as fall-run. The hatcheries upstream of the sampling site do not produce spring-run Chinook; therefore, it is assumed that catch of hatchery origin spring-run Chinook are of fall-, late fall- or winter-run Chinook which have fork lengths that overlap with natural origin spring-run Chinook.

A total of 11 natural origin steelhead was captured by the Knights Landing RSTs during the sampling season. These fish were caught from weeks 3 through 13. A total of 57 hatchery origin
steelhead was captured by the Knights Landing RSTs. These fish were caught from week 2 through week 15.

Environmental data collected at the sampling site included: river flow volume, water temperature, water transparency, and water turbidity. Sacramento River discharge was recorded at each trap check measured by the California Data Exchange Center (CDEC) Wilkins Slough gauge. These data were averaged over the calendar week for reporting. River flows at the start of the sampling season (week 39) had a weekly mean of 6,104 cubic feet per second (cfs). River flows at the end of the sampling season (week 25) had a weekly mean of $3,645 \mathrm{cfs}$. Flows varied throughout the sampling season. In week 11, weekly mean flows peaked at 26,170 cfs and the lowest weekly mean flows of 2,910 cfs was observed in week 19. Weekly mean water temperature at the start of the survey period was $18.7^{\circ} \mathrm{C}$. Temperatures varied throughout the survey period with a low weekly mean temperature of $7.2^{\circ} \mathrm{C}$ (week 53 ) and a high mean temperature of $23.7^{\circ} \mathrm{C}$ (week 23). Mean weekly water transparency varied between a high of 204.8 centimeters ( cm ) during week 48 to a low of 7.3 cm during week 11 . Mean weekly turbidity at the sampling site varied from a low of 3.2 nephelometric turbidity units (NTU) during week 48 to a high of 200.3 NTU during week 3 .

## INTRODUCTION

The purpose of the Middle Sacramento River Juvenile Salmonid Emigration Monitoring Program is to develop information on the temporal distribution, relative abundance and run composition of juvenile Chinook salmon (Oncorhynchus tshawytscha) and steelhead trout (O. mykiss) emigrating from the upper Sacramento River to the Delta. The upper Sacramento River and its tributaries provide spawning and rearing habitat for four native runs of Chinook salmon: Sacramento River winter-run (Federal and State listed endangered), Central Valley spring-run (Federal and State listed threatened), Central Valley late fall-run and Central Valley fall-run, as well as native Central Valley steelhead trout (Federal listed threatened). The monitoring program consists of two sampling locations, one near the Tisdale Weir at rkm 196 and one located 0.8 km downstream of Knights Landing, CA, at rkm 144. Data collected on the annual timing, composition and abundance of Sacramento River salmonids observed at the Tisdale Weir sampling location is detailed in a separate document. The Knights Landing sampling site is the most downstream monitoring site on the Sacramento River above the confluence with large salmonid bearing tributaries, specifically the American and Feather Rivers located at Sacramento River 96.7 rkm and 128.8 rkm, respectively. All salmonids captured by the RSTs at Knights Landing are assumed to be produced in the upper Sacramento River and its tributaries including the CNFH and LSNFH (Figure 1).

Juvenile Chinook salmon emigrate from the upper Sacramento River and its tributaries toward the Delta in a wide range of life stages (Healey 1991). Juvenile fall-run Chinook salmon have a residency period of one to seven months and typically migrate January through May. Juvenile spring-run Chinook salmon have a longer period of stream residency, between three and fifteen months, and may migrate as recently emerged fry, rear for a short period and emigrate as smolts, or rear for longer periods and emigrate as yearlings. Young-of-year (YOY) spring-run migrate between the months of March and June and between November and April as yearlings. Winter-run juveniles have a residency period of five to ten months and will migrate as YOY fry, smolts or as yearlings during the months of November through May. Juvenile late fall-run Chinook salmon may also migrate as emerged fry, smolts or yearlings and typically migrate during the months of November through May (Fisher 1994; Yoshiyama et al. 1998).

Adult Central Valley steelhead trout generally enter the Delta August through October and spawn December through April. Adult migration and spawning timing may be highly variable depending on river flows and water temperatures during migration periods. Juveniles may rear in their natal stream or affiliated tributary stream for 1-3 years. Juveniles may emigrate anywhere between 1-3 years of age, but generally leave for the ocean at 2 years of age (Hallock 1989). Emigration timing of juveniles may be highly variable and may occur at any time of the year. However, most juveniles emigrate during spring months with a smaller emigration occurring during fall months.

Two federal fish hatcheries, CNFH and LSNFH (substation of CNFH), located upstream from the sampling location, collectively produce winter-, fall- and late fall-runs of Chinook salmon, as well as Central Valley steelhead trout. These fish help supplement the natural origin populations. Prior to releasing fish into the Sacramento River, these hatcheries externally mark
all of their steelhead production and one quarter of the Chinook salmon production by removing the adipose fin. Externally marked Chinook are also given a coded wire tag (CWT). A small percentage of these hatchery released fish were captured by the RST's.

The abundance of native, anadromous salmonids in California's Central Valley has dropped precipitously because of anthropogenic changes to the environment. Loss of spawning and rearing habitat for Central Valley salmonids coupled with environmental alterations along migration corridors has put great strain on natural populations.

Much of the historic spawning habitat for Central Valley salmonids is no longer accessible. Construction of dams on many of the major salmonid bearing streams from the mid-1800's through mid-1900's blocked access to over 72\% of salmonid holding, spawning, and rearing areas (Yoshiyama et al. 2001). Dams can create unsuitable habitat downstream of the impoundment by increasing river temperatures and increasing river channelization while reducing natural river flows, natural cover and natural gravel recruitment necessary for successful spawning.

Rivers in the Central Valley have also been altered and channelized with levees to aid in flood protection of urban areas and assist in agricultural water needs. These agricultural activities may further compromise water quality with urban and agricultural runoff which often contains pollutants such as pesticides, fertilizers, and treated effluent. Increases in water turbidity from such contaminants can increase water temperatures which affect juvenile survival (Brandes and Mclain 2001; Moyle 2002). Loss of suitable rearing habitat reduces juvenile survivability during emigration which results in a reduction in the salmon population.

The demand for diverted water and associated water transfer activities in the California Central Valley alter aquatic ecosystems by creating unnatural river flow regimes, altering flow magnitude and reducing available habitat. Unscreened water diversions in migration corridors may directly impact juvenile salmonids through entrainment mortality. Entrainment of juvenile salmonids may occur at screened water diversions as well; two such diversions are the Harvey Banks Delta Pumping Plant (SWP) and the C.W. Bill Jones Pumping Plant (CVP) (Kimmerer 2008).

The altered aquatic environment in the Central Valley may promote the success of non-native fish species. Non-native fishes can negatively affect native species through competition, predation, disrupting food webs, reshaping ecosystem functions, introducing disease, or displacing native species (Mount et al. 2012). The introduction of highly efficient piscivores such as the smallmouth bass (Micropterus dolomieu), largemouth bass (M. salmoides), and striped bass (Morone saxatilis) into the Delta in the late 1800's (Dill 1997) has had considerable impacts upon native salmonid stocks. These non-native fish have been observed to forage on native salmonids at greater rates than even the largest native piscivore, the Sacramento pikeminnow (Ptychocheilus grandis) (Nobriga and Feyer 2007). Non-native piscivores occur in nearly all habitats used by emigrating and rearing salmonid juveniles, including spawning grounds in the Sacramento River, its tributaries and the Delta.

Protecting juvenile salmonids as they emigrate from their natal waters toward the Delta and onward to the Pacific Ocean is essential to maintain the existence of the remaining salmonid stocks in the Central Valley. Various restrictions have been placed upon water diversion projects within the Delta to protect juveniles during peak emigration periods. Having a near real-time estimate of abundance and emigration timing for protected salmonid species improves the ability to implement and adapt protective measures, enhancing overall protection of salmonids while augmenting water management practice flexibility.

NMFS recognized SWP and CVP Delta water operations practices to be hazardous to listed salmonid species by identifying loss at the south Delta pumping facilities or migratory delay and fish disorientation in the interior Delta. NMFS suggested Reasonable and Prudent Alternatives that would enable water export activities to continue in compliance with the Federal Endangered Species Act, including adaptive operations of the Delta Cross Channel (DCC) gates to decrease potential entrainment into the interior Delta (NMFS 2009).

CVP/SWP operations under the 2009 NMFS Operations Criteria and Procedures biological opinion (BO) rely on data collected by the California Department of Fish and Wildlife (CDFW) Middle Sacramento River Juvenile Salmonid Emigration Monitoring Program (Program) near Knights Landing to inform DCC gate operations. Additionally, monitoring data from Knights Landing are used to identify and relay to water managers emigration trends and approximate numbers of juvenile salmonids entering the Delta. Data collected by the Program were distributed to constituents by CDFW on a per-trap-check basis; the traps were serviced, data were gathered, and data were summarized in an electronic format and then distributed via email the same day.

The primary goals of the Knights Landing program are:

1. Provide early warning of emigrating listed salmonids moving toward the Delta so the CVP and SWP projects can modify their water export activities, including DCC gate closures for up to three days.
2. Document passage of emigrating salmonids including timing, relative abundance, and environmental conditions.
3. Estimate emigrating salmonid numbers in the middle Sacramento River above the Delta.
4. Develop a long-term dataset on emigration with which to compare changes over time.


Figure 1. Map of the upper Sacramento River and tributaries depicting locations of the CDFW juvenile monitoring sites, the Delta Cross Chanel Gates and the C.W. Bill Jones (Tracy) pumping facility.

## METHODS

Juvenile salmonid emigration monitoring at the Knights Landing sampling site began on September 16, 2015, and concluded on June 20, 2016, for a total of 280 days of continuous sampling. RSTs were used for sampling as they allow for data to be collected on juvenile salmonid presence and passage over time, age and size at emigration, emigration timing, and species and run composition. A detailed description of RST use and operation is described in Kennen et al. (1994) and Volkhardt et al. (2007).

The Program outfitted two RSTs with 8-foot diameter cones secured to one another and anchored in place on the east side of the Sacramento River channel (river left). The channel position of the RSTs fluctuated slightly based on Sacramento River flow. During baseflow conditions, the RSTs were positioned in the thalweg approximately 10 m from the east bank. During high flow conditions the RSTs were within approximately 3.4 m of the east bank.

Servicing of the RSTs was completed in accordance with a condition dependent sampling schedule which is an approach where environmental conditions dictate trap operation. Daily trap checks were the baseline approach to sampling under normal conditions where river flows were stable (less than $10,000 \mathrm{cfs}$ ) and in-river debris was minimal. As river conditions changed or an increase in catch was observed, various trap servicing and configuration methods were employed. (Appendix A)

Personnel accessed the RSTs using CDFW vessels which were moored on the Sacramento River at Knights Landing. These vessels included a 30' pontoon boat and a 19' Design Concepts Delta Angler. Both were outfitted with the equipment necessary to collect data and maintain the RSTs.

During each trap servicing, crews collected data specific to the performance of each RST including time since last RST service, average cone revolutions per minute, total cone revolutions since last RST service, total hours sampled, water velocity entering each RST cone, and depth of water where the RSTs were positioned. Water velocity was evaluated using a Global Water flow probe (model FP111), and water depth at each trap was estimated using a handheld electronic depth finder.

Environmental data collected and recorded during each RST service included: water temperature, water transparency, water turbidity, and river discharge volume. Water temperature was recorded over time using an electronic Onset HOBO temperature logger and during each trap service with a handheld H-B USA standard liquid thermometer. Water transparency at the sampling location was measured during each trap service using a Secchi disc following standard protocols (Orth 1983). Water turbidity was measured by collecting two water samples during each trap service and analyzed using an HF Scientific DRT-15CE turbidimeter, then averaged and reported in NTUs. River discharge volume, measured in cubic feet per second (cfs), was obtained from the California Data Exchange Center (CDEC 2016) gauge at Wilkins Slough, which is located upstream from the town of Knights Landing. River flow was an important factor for the program to consider as river flows are known to influence
juvenile emigration patterns and may create hazardous working conditions for personnel working on the traps.

All fishes captured in the RSTs were identified to species and measured to the nearest millimeter ( mm ). Salmonids greater than $40-\mathrm{mm}$ fork length ( FL ) were weighed to the nearest tenth of a gram. Run was assigned to juvenile Chinook based on FL using the LAD run identification tables (Greene 1992). Life stages were assigned based on visual appearance and recorded as alevin, fry, parr, silvery parr, or smolt. Steelhead life stage was estimated based on FL measurements. Fish measuring < 100 mm were assigned to the YOY age class, fish measuring 100 mm to 300 mm were yearlings, and fish over 300 mm were adults. Catch per unit of effort (CPUE) for each run of Chinook salmon and steelhead trout was evaluated by calculating total number of fish captured divided by the total hours of sampling. Non-salmonids were measured to total length (TL), no weights were recorded. For reporting purposes, all salmonids possessing an intact adipose fin (unmarked) were assumed to be of natural origin. It is recognized that portions of hatchery production releases contain unmarked and untagged juvenile Chinook; however, identifying them against their natural origin counterparts is not possible without genetic or otolith data analysis.

Up to 20 adipose fin-clipped, hatchery origin Chinook salmon of each run per trap maintenance event were collected. The absence of the adipose fin indicates the presence of a CWT identifying the hatchery of origin, release date, release location, and release group size. These fish were taken to a CDFW laboratory for removal of the CWT. The CWTs were read by CDFW staff and cross referenced with information provided by the federal hatcheries.

All data were recorded on water-proof datasheets, transported to the CDFW Region 2 Headquarters office, and checked for quality assurance and quality control (QAQC). Data summaries were e-mailed to resource agencies and various stakeholders on the same day to provide real-time reporting of trap catch data. Following the initial data quality check, data were entered into the Comprehensive Assessment \& Monitoring Program (CAMP) database platform developed by the United States Fish and Wildlife Service (USFWS) for analysis and reporting. Following database entry, data were again verified for QAQC using standardized protocols.

In this report, Chinook salmon and steelhead trout data were combined into weekly sums to evaluate trends in salmonid emigration timing and abundance, and to help in normalizing variation in effort and trap efficiency trials. Sample weeks began on a Sunday and ended on a Saturday, and each week of the year was assigned a number in accordance with the Julian calendar.

Trap efficiency was evaluated using mark-and-recapture methods (Volkhardt 2007). Groups of juvenile Chinook were marked externally using either Visible Implant Elastomer (VIE) tags or biological stain. Juvenile Chinook externally marked with VIE tags were first collected from CNFH. A minimum of 1000 fish were obtained and transported to our tagging station to be marked. Fish marked with the biological stain were sourced from the sampling location. When

RST catch of juvenile Chinook salmon was sufficient, a minimum of 150 fish were externally marked using Bismarck Brown (BBY) biological stain. Trap efficiency release groups were held overnight near the sampling location to assess mortality. Upon release, groups were distributed across the river channel in small groups. The release site, approximately 2 miles upstream, was selected as it was assumed that marked fish would evenly distribute across the channel and have an equal chance of being captured again by the RSTs, but not too far upstream that predation on marked fish would impact the efficiency trials.

Passage estimates were generated for Chinook salmon using the functions embedded in the RST data management and access platform developed by the USFWS CAMP. The CAMP RST platform estimates daily passage by dividing daily catch by a daily estimate of efficiency derived from efficiency trials conducted during the season. Daily catch is expanded during times where no sampling was conducted or where the half cone fishing configuration was utilized. To estimate passage during times where no sampling was conducted, the platform smooths observed CPUE through time, similar to a moving average. The CPUE is then multiplied by the number of hours the trap was not operational during the 24-hour period to estimate catch for that day. To expand catch during times where the half cone sampling configuration was utilized, daily catch was doubled as it is assumed that modifying the trap to half cone fishing reduces effort by half. To estimate efficiency every day of the season, the Platform utilizes a b-spline smoothing method to model daily efficiency. Steelhead trout life history creates uncertainty when applying trap capture efficiencies to estimate passage and passages estimates were not produced for steelhead trout.

## RESULTS

## Environmental Conditions

Mean daily flow reported at the CDEC Wilkins Slough gauge during the sampling season (September 16, 2015, through June 20, 2016) was $9,340 \mathrm{cfs}(6,636 \mathrm{cfs}$ standard deviation (SD)). Maximum flow volume recorded was 27,100 cfs during week 11 on March 14, and minimum flow volume recorded was 2,910 cfs during week 19 on May 5. (Figure 2, Table 1)

During the 2015/2016 sampling season there were three distinct flow events which varied in magnitude and duration. The first of these flow events began during week 49 when flows increased from a weekly average of 4,571 cfs to 13,633 cfs during week 52 . After declining to a weekly average of 5,190 cfs during week 53, flows again increased to average of 24,105 cfs during week 4 . Weekly flow averaged above 10,000 cfs through week 6 . The final flow event began during week 9 when average weekly flows increased from 7,397 cfs to 26,343 cfs during week 11. Average weekly flows remained below 10,000 cfs from week 15 through the end of sampling in week 25. Combined Chinook salmon catch during these events were $378,14,893$, and 1,773 , respectively, making up $94 \%$ of the season's total juvenile Chinook salmon catch.

Water temperatures generally decreased from the start of sampling efforts during week 40 through week 1 , then generally increased through the end of the sampling season. Mean water temperature during the sampling period was $14.8^{\circ} \mathrm{C}\left(4.9^{\circ} \mathrm{C}\right.$ SD). The minimum water temperature observed was $6.7^{\circ} \mathrm{C}$ recorded during week 1 on January 3 , while the maximum water temperature of $25.6^{\circ} \mathrm{C}$ was recorded on June 1, week 23. (Figure 2, Table 1)


Figure 2. Daily water temperature ( $\mathrm{C}^{\circ}$ ) values collected at the sampling site between September 16, 2015, and June 20, 2016. Water flow rate was recorded by CDEC, Wilkins Sough gauge and reported in cubic feet per second (cfs).

The minimum water transparency recorded at the sampling site was 1.3 cm during week 4 on January 20. The maximum water transparency recorded was 262.9 cm recorded during week 49 on November 30. Mean water transparency for the sampling season was 74.8 cm ( 51.7 cm SD). (Figure 3, Table 1)

Turbidity at the sampling site varied from a low of 2.6 NTU recorded during week 48 on November 28 to a high of 467 NTU recorded during week 4 on January 20. Mean turbidity during the sampling season was 34.1 NTU (45.1 NTU SD) (Figure 3, Table 1).


Figure 3. Daily water transparency (cm) and turbidity (NTU) values collected the sampling site between September 16, 2015, and June 20, 2016.

Table 1. Weekly summaries of environmental conditions recorded at the rotary screw traps located on the Sacramento River near Knights Landing, California, from September 16, 2015, through June 20, 2016.

| Week | Beginning <br> Date | Mean Water <br> Temperature <br> $\left(\mathbf{C}^{\circ}\right)$ | Mean River <br> Flow (cfs) | Mean Secchi <br> Depth (cm) | Mean Water <br> Turbidity (NTU) |
| ---: | :---: | :---: | :---: | :---: | :---: |
| 38 | $9 / 14 / 2015$ | 18.7 | 6,104 | 104.5 | 6.6 |
| 39 | $9 / 21 / 2015$ | 20.1 | 6,003 | 118.1 | 6.9 |
| 40 | $9 / 28 / 2015$ | 18.6 | 5,910 | 134.2 | 5.0 |
| 41 | $10 / 5 / 2015$ | 19.3 | 5,770 | 138.0 | 5.4 |
| 42 | $10 / 12 / 2015$ | 18.5 | 5,690 | 126.8 | 5.5 |
| 43 | $10 / 19 / 2015$ | 16.6 | 5,071 | 140.4 | 5.3 |
| 44 | $10 / 26 / 2015$ | 15.8 | 4,604 | 154.6 | 4.3 |
| 45 | $11 / 2 / 2015$ | 12.8 | 4,224 | 158.9 | 4.2 |
| 46 | $11 / 9 / 2015$ | 11.7 | 4,299 | 163.8 | 3.7 |
| 47 | $11 / 16 / 2015$ | 12.3 | 3,969 | 164.4 | 4.2 |
| 48 | $11 / 23 / 2015$ | 9.0 | 4,060 | 204.8 | 3.2 |
| 49 | $11 / 30 / 2015$ | 9.8 | 4,571 | 152.4 | 5.0 |
| 50 | $12 / 7 / 2015$ | 10.9 | 6,709 | 73.9 | 16.1 |
| 51 | $12 / 14 / 2015$ | 9.0 | 9,599 | 49.2 | 27.2 |
| 52 | $12 / 21 / 2015$ | 7.5 | 9405 | 41.5 | 62.5 |
| 53 | $12 / 28 / 2015$ | 7.2 | 5,190 | 68.6 | 15.4 |
| 1 | $1 / 4 / 2016$ | 7.5 | 7,313 | 88.1 | 20.8 |
| 2 | $1 / 11 / 2016$ | 8.6 | 13,466 | 28.8 | 60.4 |


| Week | Beginning <br> Date | Mean Water <br> Temperature <br> $\left(\mathbf{C}^{\circ}\right)$ | Mean River <br> Flow (cfs) | Mean Secchi <br> Depth (cm) | Mean Water <br> Turbidity (NTU) |
| ---: | :---: | :---: | :---: | :---: | :---: |
| 3 | $1 / 18 / 2016$ | 10.1 | 23,286 | 11.5 | 200.3 |
| 4 | $1 / 25 / 2016$ | 10.9 | 23,764 | 11.8 | 131.7 |
| 5 | $2 / 1 / 2016$ | 9.6 | 19,664 | 17.9 | 93.9 |
| 6 | $2 / 8 / 2016$ | 10.4 | 11,341 | 37.4 | 30.9 |
| 7 | $2 / 15 / 2016$ | 13.4 | 8,557 | 45.9 | 22.1 |
| 8 | $2 / 22 / 2016$ | 12.5 | 9,244 | 53.3 | 16.8 |
| 9 | $2 / 29 / 2016$ | 13.6 | 7,240 | 76.6 | 10.6 |
| 10 | $3 / 7 / 2016$ | 12.9 | 17,291 | 35.5 | 86.3 |
| 11 | $3 / 14 / 2016$ | 11.4 | 26,171 | 7.3 | 171.6 |
| 12 | $3 / 21 / 2016$ | 12.8 | 25,214 | 9.2 | 75.4 |
| 13 | $3 / 28 / 2016$ | 12.9 | 20,200 | 23.3 | 37.1 |
| 14 | $4 / 4 / 2016$ | 16.6 | 11,786 | 31.5 | 31.8 |
| 15 | $4 / 11 / 2016$ | 17.9 | 9,596 | 56.2 | 17.1 |
| 16 | $4 / 18 / 2016$ | 18.5 | 7,943 | 57.7 | 14.9 |
| 17 | $4 / 25 / 2016$ | 18.7 | 5,440 | 58.6 | 15.5 |
| 18 | $5 / 2 / 2016$ | 20.4 | 3,404 | 55.2 | 15.3 |
| 19 | $5 / 9 / 2016$ | 20.8 | 4,164 | 34.2 | 23.0 |
| 20 | $5 / 16 / 2016$ | 22.8 | 3,601 | 49.8 | 14.6 |
| 21 | $5 / 23 / 2016$ | 21.7 | 3,610 | 60.1 | 12.8 |
| 22 | $5 / 30 / 2016$ | 23.7 | 3,528 | 58.6 | 12.3 |
| 24 | $6 / 6 / 2016$ | 22.5 | 3,395 | 69.4 | 8.5 |
| 25 | $6 / 20 / 2016$ | 21.7 | 3,645 | 63.4 | 12.8 |

## Summary of Chinook Salmon Emigration

All runs and juvenile life stages of Chinook salmon were represented in the RST catch during the sampling season. A total of 18,214 juvenile salmon was captured, of which 17,745 unmarked (adipose intact) Chinook salmon accounted for $97.5 \%$ of total catch. Unmarked Chinook salmon include naturally spawned winter-run, spring-run, fall-run and late fall-run. Marked Chinook salmon catch totaled 469 , or $2.5 \%$ of total catch.

The first and last juvenile Chinook salmon were caught during week 39 , on September 24, and week 20, on May 16, respectively. Peak catch occurred during weeks 2 through 4 where 13,948 Chinook salmon, approximately $79 \%$ of the season total catch, were captured over 502 hours of monitoring.

## Natural origin Chinook Salmon

## Winter-run Chinook

All unmarked winter-run Chinook salmon were assumed to be natural origin as all upstream releases of hatchery origin Chinook were externally marked by the removal of the adipose fin prior to release. A total of 51 naturally produced winter-run Chinook salmon was caught by the

RSTs. The first fish of this run was caught during week 39, on September 24. Winter-run were consistently present in the RSTs during week 39 through week 10. Peaks in catch were observed during week $52(\mathrm{n}=14)$ and week $2(\mathrm{n}=10)$ with 14 winter-run sized fish accounting for approximately $47 \%$ of the season total catch of this run and a CPUE of 0.09 . All winter-run captured during the sampling period were BY 2015 based on their size at capture. (Table 2)

Table 2. Summary of the weekly catch of natural origin juvenile winter-run Chinook salmon from September 16, 2015, through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this run.
$\left.\begin{array}{cccccccc}\hline \text { Week } & \begin{array}{c}\text { Beginning } \\ \text { Date }\end{array} & \begin{array}{c}\text { Effort } \\ \text { (h) }\end{array} & \begin{array}{c}\text { Total } \\ \text { Catch }\end{array} & \text { CPUE } & \begin{array}{c}\text { Mean } \\ \text { FL } \\ (\mathbf{m m})\end{array} & \begin{array}{c}\text { Minimum } \\ \text { FL }(\mathbf{m m})\end{array} & \begin{array}{c}\text { Maximum } \\ \text { FL ( } \mathbf{m m})\end{array}\end{array} \begin{array}{c}\text { Standard } \\ \text { Deviation }\end{array}\right]$

## Spring-run Chinook

A total of 926 unmarked spring-run Chinook salmon was caught by the RSTs. The first spring-run sized fish was caught during week 50 on December 15. Spring-run emigration timing appeared bimodal with two peaks in catch occurring during weeks 52 through 4 ( $\mathrm{n}=102$ ) and weeks 13 through 16 ( $\mathrm{n}=746$ ). All juvenile spring-run Chinook salmon sampled by the RSTs were BY 2015 based on size at capture. (Table 3)

Table 3. Summary of the weekly catch of natural origin juvenile spring-run Chinook salmon from September 16, 2015, through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this run.

| Week | Beginning <br> Date | Effort <br> $\mathbf{( h )}$ | Total <br> Catch | CPUE | Mean <br> FL $(\mathbf{m m})$ | Minimum <br> FL $(\mathbf{m m})$ | Maximum <br> FL $(\mathbf{m m})$ | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 50 | $12 / 7 / 2015$ | 153.6 | 4 | 0.026 | 37 | 36 | 38 | 1.0 |
| 51 | $12 / 14 / 2015$ | 129.4 | 6 | 0.046 | 38 | 37 | 39 | 0.8 |
| 52 | $12 / 21 / 2015$ | 150.7 | 58 | 0.385 | 40 | 39 | 43 | 1.1 |
| 53 | $12 / 28 / 2015$ | 59.6 | 1 | 0.017 | 40 | 40 | 40 | - |
| 2 | $1 / 11 / 2016$ | 218.7 | 33 | 0.151 | 45 | 43 | 51 | 2.0 |
| 3 | $1 / 18 / 2016$ | 139.6 | 21 | 0.150 | 49 | 45 | 60 | 4.4 |


| Week | Beginning <br> Date | Effort <br> $\mathbf{( h )}$ | Total <br> Catch | CPUE | Mean <br> FL $(\mathbf{m m})$ | Minimum <br> FL $(\mathbf{m m})$ | Maximum <br> FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | $1 / 25 / 2016$ | 143.7 | 6 | 0.042 | 50 | 47 | 57 | 3.9 |
| 5 | $2 / 1 / 2016$ | 83.2 | 1 | 0.012 | 66 | 66 | 66 | - |
| 7 | $2 / 15 / 2016$ | 277.0 | 1 | 0.004 | 65 | 65 | 65 | - |
| 8 | $2 / 22 / 2016$ | 304.5 | 1 | 0.003 | 56 | 56 | 56 | - |
| 9 | $2 / 29 / 2016$ | 295.3 | 2 | 0.007 | 61 | 59 | 62 | 2.1 |
| 10 | $3 / 7 / 2016$ | 201.9 | 4 | 0.020 | 64 | 62 | 65 | 1.3 |
| 11 | $3 / 14 / 2016$ | 124.5 | 6 | 0.048 | 69 | 65 | 79 | 5.2 |
| 12 | $3 / 21 / 2016$ | 183.2 | 26 | 0.142 | 77 | 67 | 93 | 7.7 |
| 13 | $3 / 28 / 2016$ | 298.5 | 207 | 0.693 | 77 | 70 | 92 | 3.9 |
| 14 | $4 / 4 / 2016$ | 238.4 | 246 | 1.032 | 78 | 73 | 89 | 3.0 |
| 15 | $4 / 11 / 2016$ | 274.8 | 122 | 0.444 | 84 | 77 | 99 | 4.0 |
| 16 | $4 / 18 / 2016$ | 295.2 | 171 | 0.579 | 84 | 80 | 95 | 3.0 |
| 17 | $4 / 25 / 2016$ | 329.9 | 7 | 0.021 | 91 | 85 | 99 | 5.0 |
| 18 | $5 / 2 / 2016$ | 325.2 | 2 | 0.006 | 90 | 90 | 90 | 0.0 |
| 19 | $5 / 9 / 2016$ | 270.7 | 1 | 0.004 | 96 | 96 | 96 | - |

Fall-run Chinook
A total of 16,767 unmarked fall-run Chinook salmon was caught by the RSTs. The first fall-run was caught during week 50, on December 15, and fall-run Chinook were present throughout the remainder of the survey period with few exceptions. Catch peaked during weeks 2 through 5 with a total of 14,702 fall-run captured, representing approximately $88 \%$ of total natural origin fall-run Chinook catch. All juvenile fall-run Chinook salmon sampled by the RSTs were BY 2015 based on size at capture. (Table 4)

Table 4. Summary of the weekly catch of natural origin juvenile fall-run Chinook salmon from September 16, 2015, through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this run.

| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $(\mathbf{m m})$ | Minimum <br> FL (mm) | Maximum <br> FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 50 | $12 / 7 / 2015$ | 153.6 | 1 | 0.007 | 33 | 33 | 33 | - |
| 51 | $12 / 14 / 2015$ | 129.4 | 3 | 0.023 | 35 | 34 | 36 | 1.0 |
| 52 | $12 / 21 / 2015$ | 150.7 | 237 | 1.573 | 36 | 30 | 39 | 1.5 |
| 53 | $12 / 28 / 2015$ | 59.6 | 11 | 0.185 | 36 | 35 | 39 | 1.4 |
| 1 | $1 / 4 / 2016$ | 206.6 | 9 | 0.044 | 38 | 36 | 41 | 1.5 |
| 2 | $1 / 11 / 2016$ | 218.7 | 4306 | 19.689 | 37 | 33 | 43 | 1.8 |
| 3 | $1 / 18 / 2016$ | 139.6 | 5977 | 42.815 | 37 | 31 | 45 | 1.9 |
| 4 | $1 / 25 / 2016$ | 143.7 | 3584 | 24.941 | 38 | 29 | 46 | 2.0 |
| 5 | $2 / 1 / 2016$ | 83.2 | 835 | 10.036 | 38 | 31 | 48 | 2.5 |
| 6 | $2 / 8 / 2016$ | 209.3 | 74 | 0.354 | 39 | 34 | 52 | 3.4 |
| 7 | $2 / 15 / 2016$ | 277.0 | 30 | 0.108 | 44 | 35 | 54 | 5.8 |


| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $(\mathbf{m m})$ | Minimum <br> FL (mm) | Maximum <br> FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 8 | $2 / 22 / 2016$ | 304.5 | 27 | 0.089 | 45 | 25 | 55 | 7.1 |
| 9 | $2 / 29 / 2016$ | 295.3 | 21 | 0.071 | 48 | 39 | 57 | 5.4 |
| 10 | $3 / 7 / 2016$ | 201.9 | 86 | 0.426 | 41 | 32 | 62 | 7.9 |
| 11 | $3 / 14 / 2016$ | 124.5 | 221 | 1.775 | 44 | 31 | 65 | 8.4 |
| 12 | $3 / 21 / 2016$ | 183.2 | 191 | 1.043 | 45 | 26 | 69 | 9.8 |
| 13 | $3 / 28 / 2016$ | 298.5 | 194 | 0.650 | 55 | 33 | 72 | 12.8 |
| 14 | $4 / 4 / 2016$ | 238.4 | 128 | 0.537 | 66 | 38 | 75 | 8.7 |
| 15 | $4 / 11 / 2016$ | 274.8 | 123 | 0.448 | 72 | 44 | 79 | 6.4 |
| 16 | $4 / 18 / 2016$ | 295.2 | 542 | 1.836 | 75 | 55 | 82 | 4.0 |
| 17 | $4 / 25 / 2016$ | 329.9 | 21 | 0.064 | 75 | 61 | 85 | 7.3 |
| 18 | $5 / 2 / 2016$ | 325.2 | 11 | 0.034 | 84 | 80 | 89 | 3.1 |
| 19 | $5 / 9 / 2016$ | 270.7 | 39 | 0.144 | 81 | 66 | 94 | 7.0 |
| 20 | $5 / 16 / 2016$ | 274.3 | 5 | 0.018 | 81 | 75 | 86 | 4.2 |

Late fall-run Chinook
A total of 1 unmarked late fall-run Chinook salmon was captured during week 52 on December 24. All juvenile late fall-run Chinook salmon sampled by the RSTs were BY 2015 based on size at capture. (Table 5)

Table 5. Summary of the weekly catch of natural origin juvenile late fall-run Chinook salmon from September 16, 2015, through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this run.

| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $(\mathrm{mm})$ | Minimum <br> FL $(\mathrm{mm})$ | Maximum <br> FL $(\mathrm{mm})$ | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 52 | $12 / 21 / 2015$ | 150.7 | 1 | 0.007 | 120 | 120 | 120 | - |

## Hatchery Origin Chinook Salmon

Upstream production releases from CNFH and LSNFH consisted only of winter-run ( $n=419,690$ ), fall-run ( $n=12,160,858$ ) and late fall-run ( $n=474,938$ ) Chinook. It is the intention of both hatcheries to mark, by the removal the adipose fin, and tag, with a CWT, at least 25 percent of hatchery origin fish under the guidelines of the Constant Fractional Marking Program (PalmerZwahlen et al 2019). Hatchery origin winter-run and late fall-run Chinook salmon are to be $100 \%$ marked. However, due to error associated with the marking and tagging equipment utilized in this process, portions of each release were not marked and/or tagged. Of the total winter-run Chinook salmon released by LSNFH, 2,638 (0.6\%) were marked but not tagged, 941 ( $0.2 \%$ ) were tagged but not marked, and 246 ( $0.1 \%$ ) were not marked or tagged. The remainder were marked and tagged ( $\mathrm{n}=415,865,99.1 \%$ ). Of the total late fall-run Chinook salmon released by CNFH, $1,700(0.4 \%)$ were marked but not tagged, 8,975 ( $1.9 \%$ ) were tagged but not marked, and $339(0.1 \%)$ were not marked or tagged. The remainder were marked and tagged
( $n=463,924,97.7 \%$ ). Fall-run Chinook salmon are fractionally marked at a rate of $25 \%$. Of the total fall-run Chinook released by CNFH, 2,779 (<0.1\%) were marked but not tagged, 907 (<0.1\%) tagged but not marked, 9,123,440 (75.0\%) were not marked or tagged, and 3,033,741 (24.9) were marked and tagged. (Table 6)

Following releases, 469 adipose fin-clipped Chinook salmon were subsequently captured by the RSTs consisting of all 4 runs using the LAD criteria for run determination: 52 winter-run (10.7\%), 159 spring-run (33.9\%), 203 fall-run (42.3\%), and 55 late fall-run (11.7\%).

Table 6. Summary of hatchery origin juvenile Chinook salmon and steelhead trout by CNFH and LSNFH, released upstream from the Knights Landing sampling site from September 16, 2015, through June 20, 2016.

| Run or <br> Species | Week | Release Dates | Number <br> marked <br> with CWT | Number <br> marked <br> without <br> CWT | Number <br> unmarked <br> with CWT | Number <br> Unmarked | Release <br> location $^{1}$ |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Late Fall | 50 | $12 / 9 / 2015$ | 253,957 | 923 | 6,556 | 339 | CNFH |
| Late Fall | 50 | $12 / 11 / 2015$ | 76,525 | 777 | 388 | - | CNFH |
| Late Fall | 52 | $12 / 22 / 2015$ | 67,148 | 0 | 678 | - | CNFH |
| Late Fall | 3 | $1 / 12 / 2016$ | 66,294 | 0 | 1,353 | - | CNFH |
| Winter | 8 | $2 / 17 / 2016-$ | 415,865 | 2,638 | 941 | 246 | LRP |
|  |  | $2 / 18 / 2016$ |  |  |  |  |  |
| Fall | 11 | $3 / 14 / 2016$ | 214,826 | 857 | 0 | 647,824 | CNFH |
| Fall | 13 | $3 / 22 / 2016$ | 324,065 | - | 0 | 973,032 | CNFH |
| Fall | 15 | $4 / 7 / 2016$ | $1,391,663$ | 1,179 | 232 | $4,184,565$ | CNFH |
| Fall | 16 | $4 / 12 / 2016$ | 634,026 | 734 | 675 | $1,908,277$ | CNFH |
| Fall | 18 | $4 / 29 / 2016$ | 469,161 | - | - | $1,409,742$ | CNFH |
| Steelhead | $1-2$ | $1 / 4 / 2016-$ | - | 585,127 | - | 6,235 | BB |

${ }^{1}$ LRP = Lake Redding Park; CNFH = Coleman National Fish Hatchery; BB = Bend Bridge.

## Winter-run Chinook

A total of 52 hatchery origin juvenile winter-run Chinook salmon was captured by the RSTs. The first hatchery origin winter-run based on LAD criteria was captured during week 3 which was prior to the first release of hatchery-reared winter-run Chinook. CWT data collected from these individuals confirmed these were late-fall run Chinook released from CNFH. It is likely that subsequent catch of hatchery origin LAD winter-run Chinook included individuals from the CNFH late fall-run releases. The last hatchery origin winter-run was captured during week 11. All hatchery origin winter-run Chinook were BY 2015 (Table 7).

Table 7. Summary of weekly catch of hatchery origin juvenile winter-run Chinook salmon from September 16, 2015, through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this run.

| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $(\mathbf{m m})$ | Minimum <br> FL (mm) | Maximum <br> FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | $1 / 18 / 2016$ | 139.6 | 1 | 0.007 | 88 | 88 | 88 | - |
| 8 | $2 / 22 / 2016$ | 304.5 | 24 | 0.079 | 95 | 79 | 109 | 7.8 |
| 9 | $2 / 29 / 2016$ | 295.3 | 20 | 0.068 | 97 | 80 | 109 | 7.8 |
| 10 | $3 / 7 / 2016$ | 201.9 | 6 | 0.030 | 95 | 90 | 100 | 3.9 |
| 11 | $3 / 14 / 2016$ | 124.5 | 1 | 0.008 | 93 | 93 | 93 | - |

## Spring-run Chinook

A total of 159 hatchery origin juvenile Chinook was identified as spring-run Chinook salmon using LAD methodology; however, upstream hatcheries do not produce spring-run Chinook. Based on length frequency information provided by CNFH and LSNFH, it is likely these fish were part of the winter-run production releases from LSNFH and fall-run production releases from CNFH (Table 8).

Table 8. Summary of weekly catch of hatchery origin juvenile spring-run Chinook salmon from September 16, 2015, through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this run.

| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $(\mathbf{m m})$ | Minimum <br> FL (mm) | Maximum <br> FL(mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | $2 / 29 / 2016$ | 295.3 | 1 | 0.003 | 73 | 73 | 73 | - |
| 10 | $3 / 7 / 2016$ | 201.9 | 1 | 0.005 | 76 | 76 | 76 | - |
| 13 | $3 / 28 / 2016$ | 298.5 | 23 | 0.077 | 76 | 71 | 80 | 2.4 |
| 14 | $4 / 4 / 2016$ | 238.4 | 66 | 0.277 | 79 | 74 | 85 | 2.6 |
| 15 | $4 / 11 / 2016$ | 274.8 | 35 | 0.127 | 83 | 77 | 94 | 3.8 |
| 16 | $4 / 18 / 2016$ | 295.2 | 32 | 0.108 | 84 | 80 | 92 | 3.5 |
| 19 | $5 / 9 / 2016$ | 270.7 | 1 | 0.004 | 98 | 98 | 98 | - |

## Fall-run Chinook

A total of 203 hatchery origin fall-run Chinook was captured. The first of these was captured during week 12 on March 21 following upstream releases. The last hatchery origin fall-run Chinook was captured during week 20. (Table 9)

Table 9. Summary of weekly catch of hatchery origin juvenile fall-run Chinook salmon from September 16, 2015, through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this run.

| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $\mathbf{( m m})$ | Minimum FL <br> $(\mathbf{m m})$ | Maximum <br> FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 12 | $3 / 21 / 2016$ | 183.2 | 1 | 0.005 | 64 | 64 | 64 | - |
| 13 | $3 / 28 / 2016$ | 298.5 | 3 | 0.010 | 69 | 69 | 70 | 0.6 |
| 14 | $4 / 4 / 2016$ | 238.4 | 12 | 0.050 | 72 | 68 | 74 | 1.6 |
| 15 | $4 / 11 / 2016$ | 274.8 | 25 | 0.091 | 75 | 67 | 79 | 2.9 |
| 16 | $4 / 18 / 2016$ | 295.2 | 144 | 0.488 | 75 | 64 | 81 | 3.8 |
| 17 | $4 / 25 / 2016$ | 329.9 | 4 | 0.012 | 78 | 71 | 83 | 5.1 |
| 19 | $5 / 9 / 2016$ | 270.7 | 13 | 0.048 | 80 | 72 | 90 | 5.4 |

Late Fall-run Chinook
A total of 55 hatchery origin late-fall Chinook salmon was captured. The first hatchery origin late fall-run was captured during week 50 on December 14 and the last hatchery origin late fallrun was captured during week 5, on February 2; all were BY 2015 (Table 10).

Table 10. Summary of weekly catch of hatchery origin juvenile late fall-run Chinook salmon from September 16, 2015, through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this run.

| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $(\mathbf{m m})$ | Minimum <br> FL (mm) | Maximum <br> FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 50 | $12 / 7 / 2015$ | 153.6 | 18 | 0.117 | 133 | 102 | 167 | 16.8 |
| 51 | $12 / 14 / 2015$ | 129.4 | 3 | 0.023 | 137 | 123 | 160 | 20.3 |
| 52 | $12 / 21 / 2015$ | 150.7 | 10 | 0.066 | 139 | 115 | 170 | 16.4 |
| 53 | $12 / 28 / 2015$ | 59.6 | 1 | 0.017 | 162 | 162 | 162 | - |
| 1 | $1 / 4 / 2016$ | 206.6 | 1 | 0.005 | 122 | 122 | 122 | - |
| 2 | $1 / 11 / 2016$ | 218.7 | 1 | 0.005 | 120 | 120 | 120 | - |
| 3 | $1 / 18 / 2016$ | 139.6 | 20 | 0.143 | 167 | 143 | 190 | 15.4 |
| 5 | $2 / 1 / 2016$ | 83.2 | 1 | 0.012 | 128 | 128 | 128 | - |

## Summary of Steelhead Trout Emigration

Both, natural origin and hatchery-produced steelhead were captured at Knights Landing. Like Chinook salmon, hatchery origin steelhead are identified by the absence of an adipose fin; however, a portion of the hatchery released steelhead retained their adipose fin because of error associated with the equipment that performs adipose fin removal. A total of 591,362 hatchery origin steelhead was released by CNFH, of which 6,235 ( $0.01 \%$ ) possessed an adipose fin. For the purposes of this report, any steelhead captured by the RSTs which had an intact adipose fin was assumed to be of natural origin.

A total of 68 steelhead trout was captured. Eleven of these were of natural origin and 57 were of hatchery origin. The first and last natural origin steelhead trout were caught during week 2 , on January 11, and week 15, on April 12, respectively. Peak catch occurred during week 6 where 62 steelhead trout, or $49.2 \%$ of the season's total catch, were captured over 180 hours of monitoring.

## Natural Origin Steelhead Trout

A total of 11 natural origin steelhead trout was captured by the RSTs. One was classified as a young-of-year, measuring less than 100 mm . Ten were categorized as yearlings, measuring between 100 mm and 200 mm FL. The first and last captures occurred during week 3 and week 13 , respectively. Peak catch occurred during week 13 ( $27 \%$ of total), and the peak CPUE (0.024/hour) occurred during week 5. (Table 11)

Table 11. Summary of weekly catch of natural origin juvenile steelhead trout from September 16,2015 , through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this species.

| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $(\mathbf{m m})$ | Minimum <br> FL (mm) | Maximum <br> FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | $1 / 18 / 2016$ | 139.6 | 1 | 0.007 | 226 | 226 | 226 | - |
| 4 | $1 / 25 / 2016$ | 143.7 | 2 | 0.014 | 250 | 249 | 250 | 0.7 |
| 5 | $2 / 1 / 2016$ | 83.2 | 2 | 0.024 | 208 | 206 | 209 | 2.1 |
| 6 | $2 / 9 / 2016$ | 209.3 | 1 | 0.005 | 242 | 242 | 242 | - |
| 10 | $3 / 7 / 2016$ | 201.9 | 1 | 0.005 | 262 | 262 | 262 | - |
| 11 | $3 / 14 / 2016$ | 124.5 | 1 | 0.008 | 222 | 222 | 222 | - |
| 13 | $3 / 28 / 2016$ | 298.5 | 3 | 0.010 | 180 | 55 | 244 | 108.3 |

## Hatchery Origin Steelhead Trout

A total of 57 hatchery origin steelhead trout was captured by the RSTs. Release data provided by CNFH identify these as yearling BY 2015 steelhead. (Table 12)

Table 12. Summary of weekly catch of hatchery origin juvenile steelhead trout from September 16,2015 , through June 20, 2016. Weeks during the monitoring season not presented here resulted in zero catch of this species.

| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $(\mathbf{m m})$ | Minimum <br> FL (mm) | Maximum <br> FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | $1 / 11 / 2016$ | 218.7 | 12 | 0.055 | 216 | 142 | 257 | 29.0 |
| 3 | $1 / 18 / 2016$ | 139.6 | 16 | 0.115 | 230 | 209 | 260 | 15.8 |
| 4 | $1 / 25 / 2016$ | 143.7 | 5 | 0.035 | 232 | 212 | 249 | 14.2 |
| 5 | $2 / 1 / 2016$ | 83.2 | 11 | 0.132 | 226 | 205 | 251 | 16.0 |
| 6 | $2 / 8 / 2016$ | 209.3 | 1 | 0.005 | 275 | 275 | 275 | - |
| 8 | $2 / 22 / 2016$ | 304.5 | 1 | 0.003 | 230 | 230 | 230 | - |
| 9 | $2 / 29 / 2016$ | 295.3 | 1 | 0.003 | 237 | 237 | 237 | - |


| Week | Beginning <br> Date | Effort <br> (h) | Total <br> Catch | CPUE | Mean <br> FL <br> $(\mathbf{m m})$ | Minimum <br> FL (mm) | Maximum <br> FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 10 | $3 / 7 / 2016$ | 201.9 | 1 | 0.005 | 212 | 212 | 212 | - |
| 11 | $3 / 14 / 2016$ | 124.5 | 2 | 0.016 | 232 | 210 | 254 | 31.1 |
| 12 | $3 / 21 / 2016$ | 183.2 | 4 | 0.022 | 231 | 211 | 260 | 23.7 |
| 13 | $3 / 28 / 2016$ | 298.5 | 1 | 0.003 | 211 | 211 | 211 | - |
| 14 | $4 / 4 / 2016$ | 238.4 | 1 | 0.004 | 310 | 310 | 310 | - |
| 15 | $4 / 11 / 2016$ | 274.8 | 1 | 0.004 | 250 | 250 | 250 | - |

## Trap Efficiency Trials and Passage Estimates

A total of 11,029 juvenile Chinook salmon was marked and used in 12 efficiency trials in 2016. The mean efficiency for the season was $0.52 \%$. During the trials, salmon were marked externally using either BBY stain, or VIE tags of a specified color: pink, blue, orange, or purple. BBY staining was used on a total of 6,485 fish during weeks $2,3,4,15$, and 16 . The lowest BBY recapture rate was $0.17 \%$ during weeks 3 and 4 , and the highest recapture rate was $1.31 \%$ during week 2. VIE tags were used on a total of 4,544 fish and trials took place during weeks 6-$8,11-15$, and 17. The highest recapture rate was $1.36 \%$ during week 14 . No tagged fish were recovered during weeks 6-8, 12-13, 15, and 17. (Table 13)

Table 13. Summary of capture efficiency trials initiated from September 16, 2015, through June 20, 2016.

| Week | Dates | Mark Type | Marked <br> Released | Marked <br> Recaptured | Efficiency (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 2 | $1 / 10 / 2016-1 / 16 / 2016$ | BBY | 3202 | 43 | $1.34 \%$ |
| $3-4$ | $1 / 23 / 2016-1 / 29 / 2016$ | BBY | 2864 | 5 | $0.17 \%$ |
| 6 | $2 / 10 / 2016$ | VIE (pink) | 632 | 0 | $0.00 \%$ |
| 7 | $2 / 17 / 2016$ | VIE (blue) | 479 | 0 | $0.00 \%$ |
| 8 | $2 / 24 / 2016$ | VIE (orange) | 489 | 0 | $0.00 \%$ |
| 11 | $3 / 16 / 2016$ | VIE (purple) | 490 | 1 | $0.20 \%$ |
| 12 | $3 / 23 / 2016$ | VIE (orange) | 488 | 0 | $0.00 \%$ |
| 13 | $3 / 30 / 2016$ | VIE (purple) | 470 | 0 | $0.00 \%$ |
| 14 | $4 / 6 / 2016$ | VIE (orange) | 516 | 7 | $1.36 \%$ |
| 15 | $4 / 13 / 2016$ | VIE (pink) | 473 | 0 | $0.00 \%$ |
| $15-16$ | $4 / 16 / 2016-4 / 22 / 2016$ | BBY | 419 | 1 | $0.24 \%$ |
| 17 | $4 / 27 / 2016$ | VIE (orange) | 507 | 0 | $0.00 \%$ |

## Passage Estimates

Annual passage for each run was estimated from the beginning of the week where the first catch of that run was observed to the end of the week where the last catch of that run was observed. It is estimated that a total of 9,434,470 fall-run, 357,030 spring-run, and 11,472 winter-run Chinook salmon passed the monitoring site between September 16, 2015, and June 20, 2016. No passage estimate was calculated for late fall-run Chinook salmon as too few individuals were captured to create a reliable estimation. (Table 14)

Table 14. Estimates of natural origin Chinook salmon that passed the Knights Landing sampling location from September 16, 2015, through June 20, 2016, and associated $95 \%$ confidence interval (CI).

| Run | Passage | Lower 95\% CI | Upper 95\% CI |
| :---: | :---: | :---: | :---: |
| Fall | $8,434,470$ | $4,299,600$ | $16,482,460$ |
| Spring | 357,030 | 195,073 | 739,450 |
| Winter | 11,472 | 8,264 | 14,559 |

## Non-target Species

Non-target species include all fishes observed that were not Chinook salmon or steelhead. A total of 5,618 non-target fishes representing 34 species was captured, 10 of which are native to the Sacramento River and its tributaries and 24 are introduced species (Table 15). Some related genera catch totals were combined because juveniles have similar morphological features. For example, Pacific and river lamprey (Lampetra spp.) ammocetes were combined and totaled 180 fish. Unknown sunfish (Lepomis spp.) and unknown bass (Micropteris spp.) were combined with other unknown centrarchids and totaled 126 fish. Lastly, juvenile minnows (Cyprinidae spp.), totaling 13,530 fish, were excluded from final percentages and, instead, were collectively inventoried due to ambiguity of identifying characteristics at larval stages. The remaining 3,585 fish were comprised of $41 \%$ native fishes and $59 \%$ non-native fishes.

Table 15. Summary of non-target fish species captured between September 16, 2015, and June 20, 2016.

| Common Name | Scientific Name | Number Captured | $\begin{gathered} \hline \text { Mean FL } \\ (\mathrm{mm}) \end{gathered}$ | Min FL (mm) | Max FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Threadfin Shad | Dorosoma petenense | 1,790 | 90 | 39 | 170 | 12.7 |
| Sacramento Splittail | Pogonichthys macrolepidotus | 1,015 | 39 | 21 | 406 | 35.6 |
| Inland Silverside | Menidia beryllina | 575 | 58 | 20 | 171 | 18.4 |
| Sacramento <br> Pikeminnow | Ptychocheilus grandis | 427 | 85 | 28 | 250 | 28.2 |
| Bluegill | Lepomis macrochirus | 321 | 42 | 23 | 140 | 16.6 |
| Mosquitofish | Gambusia affinis | 232 | 31 | 18 | 67 | 6.5 |
| Golden Shiner | Notemigonus crysoleucas | 149 | 54 | 22 | 129 | 30.6 |
| Unknown Sunfish | Centrarchidae spp. | 149 | 40 | 21 | 77 | 12.7 |
| Goldfish | Carassius auratus | 130 | 48 | 22 | 400 | 41.8 |
| White Crappie | Pomoxis annularis | 89 | 46 | 20 | 245 | 32 |
| Unknown Crappie | Pomoxis spp. | 71 | 31 | 23 | 47 | 7.1 |
| Black Crappie | Pomoxis nigromaculatus | 69 | 77 | 33 | 300 | 57.9 |


| Common Name | Scientific Name | Number Captured | $\begin{gathered} \hline \text { Mean FL } \\ (\mathrm{mm}) \end{gathered}$ | Min FL (mm) | Max FL (mm) | Standard <br> Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Pacific Lamprey | Entosphenus tridentatus | 65 | 141 | 42 | 482 | 69.7 |
| Unknown Centrarchid | Centrarchidae spp. | 54 | 32 | 19 | 47 | 7.6 |
| Unknown Bass | Micropterus spp. | 53 | 23 | 12 | 38 | 4.1 |
| River Lamprey | Lampetra ayresi | 51 | 148 | 90 | 510 | 56.1 |
| Unknown Lamprey | Lampetra spp. OR Entosphenus spp. | 40 | 144 | 109 | 166 | 12.9 |
| Unknown Minnow | Cyprinidae spp. | 40 | 30 | 20 | 46 | 5.6 |
| Common Carp | Cyprinus carpio | 38 | 383 | 20 | 780 | 317.7 |
| Redear Sunfish | Lepomis microlophus | 27 | 78 | 45 | 212 | 44.4 |
| Sacramento Sucker | Catostomus occidentalis | 27 | 166 | 22 | 650 | 205.9 |
| Green Sunfish | Lepomis cyanellus | 24 | 59 | 33 | 130 | 20.6 |
| White Catfish | Ameiurus catus | 23 | 174 | 50 | 368 | 110.3 |
| Smallmouth Bass | Micropterus dolomieu | 21 | 150 | 33 | 345 | 69 |
| Red Shiner | Cyprinella lutrensis | 18 | 53 | 31 | 135 | 6.2 |
| Brown Bullhead | Ameiurus nebulosus | 13 | 178 | 48 | 432 | 106.3 |
| Warmouth | Lepomis gulosus | 13 | 55 | 35 | 85 | 11.1 |
| Striped Bass | Morone saxatilis | 12 | 305 | 40 | 520 | 158.4 |
| Three-Spined Stickleback | Gasterosteus aculeatus | 12 | 34 | 28 | 45 | 4.7 |
| American Shad | Alosa sapidissima | 11 | 98 | 30 | 342 | 90.3 |
| Channel Catfish | Ictalurus punctatus | 10 | 175 | 53 | 622 | 183.3 |
| Spotted Bass | Micropterus punctulatus | 9 | 162 | 90 | 420 | 111.2 |
| Tule Perch | Hysterocarpus traskii | 8 | 69 | 33 | 156 | 43.9 |
| Black Bullhead | Ameiurus melas | 6 | 135 | 114 | 168 | 21.7 |
| Fathead <br> Minnow | Pimephales promelas | 6 | 54 | 47 | 61 | 5.8 |
| Largemouth Bass | Micropterus salmoides | 5 | 53 | 31 | 90 | 28.4 |
| Unknown Catfish | Ictaluridae spp. | 5 | 136 | 17 | 490 | 201 |
| California Roach | Hesperoleucus symmetricus | 3 | 51 | 37 | 65 | 14 |
| Wakasagi | Hypomesus nipponensis | 3 | 57 | 43 | 75 | 16.4 |
| Hardhead | Mylopharodon conocephalus | 2 | 69 | 49 | 88 | 27.6 |


| Common Name | Scientific Name | Number <br> Captured | Mean FL <br> $(\mathrm{mm})$ | Min FL <br> $(\mathrm{mm})$ | Max FL <br> $(\mathrm{mm})$ | Standard <br> Deviation |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Hitch | Lavinia exilicauda | 2 | 114 | 98 | 130 | 22.6 |

## DISCUSSION

Several studies have suggested that increased flow, reduced water temperatures, and increases in water turbidity promote the downstream migration of juvenile salmonids (Michel et al. 2013; Kemp et al. 2005; Giorgi et al. 1997). During the 2015/2016 sampling season there were three distinct flow events which varied in magnitude and duration. While catch data resolution, sampling effort, trap capture efficiency, and uncertainty in the geographic distance fish travel prior to capture makes correlating emigration cues with catch data difficult, increases in juvenile salmonid presence were observed with each increase in flow (Figure 4), a trend that is similar to those observed in previous years.


Figure 4. Weekly average flow measured at the CDEC Wilkins Slough gauge and total weekly catch of Chinook salmon between September 16, 2015, and June 20, 2016.

An important factor affecting potential capture at the Knights Landing sampling site, and therefore passage estimates, is juvenile salmonid emigration routes. All juvenile salmonids emigrating down the Sacramento River are assumed to have the potential of being captured at the Knights Landing sampling site if they remain in the main channel from point of origin to the
sampling site. In times of excessive river flow, upstream flood control diversions, including Moulton, Colusa, and Tisdale weirs, divert Sacramento River flows and entrain juvenile salmonids in the Sutter Bypass (Figure 5). Salmonids emigrating through the Sutter Bypass are unable to return to the Sacramento River until they reach rkm 135 near the Fremont Weir which is downstream of the Knights Landing sampling site. When this occurs, observed increases in emigration can be muted at the Knights Landing sampling site affecting the ability to forecast the presence of juvenile salmonids that may be drawn into CVP and SWP facilities.


Figure 5. Map of the upper and middle Sacramento River and tributaries depicting location of the CDFW juvenile monitoring site in relation to flood relief structures.

Examining the effect of weir overtopping events on salmonid emigration monitoring may be accomplished through comparison of salmonid capture by sampling season and by monitoring location, with the caveat that other factors influencing capture must be taken into consideration, e.g., seasonal differences in juvenile production, flow, turbidity, predation, trap capture efficiency, etc. For example, a total of 18,214 Chinook salmon was caught during the 2015/2016 sampling year which is significantly less than the 2013/2014 total catch of 106,592 Chinook salmon. During the 2015/2016 sampling year, the Sacramento River flood relief weirs were overtopped 3 times resulting in a combined 59 days allowing emigrating salmonids to enter the flood plain habitats of the Sutter Bypass. Moulton Weir overtopped during week 11 for two days. Colusa and Tisdale Weirs overtopped during weeks 3 through 5 and then again during weeks 10 through 12 for a total of 25 and 34 days, respectively. An apparent decrease in trends of Chinook salmon catch was observed following the first spilling events at Colusa and Tisdale Weirs possibly indicating significant portions of the emigrating Chinook salmon population entered the Sutter Bypass. (Figure 6)


Figure 6. Daily average elevation above the weir crests at Moulton, Colusa and Tisdale weirs and total Chinook salmon catch January through March 2016, at Knights Landing on the Sacramento River.

In contrast, during the 2013/2014 season there were no overtopping events and all emigrating juvenile salmonids were restricted to the confines of the Sacramento River's main channel. The difference in catch between the 2013/2014 and 2015/2016 survey years may demonstrate the influence active flood relief weirs have on observations at monitoring locations and salmon emigration routes. This could be validated by comparisons of capture data at monitoring locations upstream of the weirs that convey flows into the Sutter Bypass (e.g., the Tisdale RST sampling location (Table 16).

Table 16. Potential for the 2015 water year overtopping events to influence salmonid capture data for natural origin and hatchery origin Chinook salmon by run and yearling steelhead trout. Natural origin late fall-run were omitted from this table because of low catch numbers ( $n=1$ ).

| ESU/Origin | Weeks | Date range | Proportion of catch following <br> overtopping events |
| :---: | :---: | :---: | :---: |
| Fall-run <br> (natural origin) | 51 to 20 | $12 / 15 / 2015$ to <br> $5 / 16 / 2016$ | $46.0 \%(n=8,641)$ |
| Spring-run <br> (natural origin) | 51 to 19 | $12 / 15 / 2015$ to <br> $5 / 12 / 2016$ | $85.9 \%(n=796)$ |
| Winter-run <br> (natural origin) | 39 to 10 | $9 / 24 / 2015$ to $3 / 9 / 2015$ | $17.6 \%(n=9)$ |
| Fall-run (hatchery) | 12 to 19 | $3 / 21 / 2016$ to <br> $5 / 15 / 2016$ | $100 \%(n=203)$ |
| Spring-run (hatchery) | 8 to 19 | $2 / 28 / 2016$ to <br> $5 / 11 / 2016$ | $100 \%(n=159)$ |
| Winter-run <br> (hatchery) | 3 to 11 | $1 / 18 / 2016$ to <br> $3 / 17 / 2016$ | $100 \%(n=52)$ |
| Late fall-run <br> (hatchery) | 51 to 5 | $12 / 14 / 2015$ to <br> $2 / 3 / 2016$ | $1.8 \%(n=1)$ |
| Steelhead <br> (natural origin) | 6 to 21 | $2 / 10 / 2015$ to <br> $5 / 25 / 2015$ | $100 \%(n=6)$ |
| Steelhead (hatchery) | 2 to 14 | $1 / 13 / 2015$ to $4 / 7 / 2015$ | $100 \%(n=120)$ |

Despite uncertainties in catch data introduced by weir overtopping events, data gathered from the sampling program does provide clear insight into juvenile salmonid migration timing and thus provides early warning of listed salmonid emigration as they move toward the Delta. Data collected during the 2015/2016 Lower Sacramento River Juvenile Salmonid Emigration Program fulfilled the program's goals of:

1. Providing early warning of emigrating listed salmonids moving into the Delta so the CVP and SWP projects could modify their water export activities, including DCC gate closures for a period sufficient to minimize entrainment of juveniles into the south Delta;
2. Documented passage of emigrating salmonids including timing, relative abundance, and response to environmental conditions;
3. Estimated emigrating salmon numbers in the lower Sacramento River above the Delta at Knights Landing;
4. Contributed to the long-term dataset on emigration which is used to compare changes over time.

As the Sutter Bypass may provide an important and needed rearing opportunity for juvenile salmonids in the Sacramento River corridor, future data collection efforts for the North Central Region's Sacramento River Juvenile Salmonid Monitoring Program should be targeted at better defining entrainment into the Sutter Bypass.

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## REFERENCES

Brandes, P. L., and J. S. McLain. 2001. Juvenile Chinook salmon abundance, distribution, and survival in the Sacramento-San Joaquin Estuary. Contributions to the Biology of the Central Valley Salmonids, Fish Bulletin 179: Volume 2. Sacramento (CA): California Department of Fish and Game. p 39-136.

California Department of Water Resources, California Data Exchange Center (CDEC), Wilkins Slough gauges. Data retrieved between 2012 and 2013 from http://cdec.water.ca.gov/

Dill, W.A., and A.J. Cordone. 1997. History and status of introduced fishes in California, 18711996. California Department of Fish and Wildlife. Fish Bulletin 178.

Fisher, F. W. 1992. Chinook salmon, Oncorhynchus tshawytscha, growth and occurrence in the Sacramento-San Joaquin River system. California Department Fish and Wildlife, Inland Fisheries Division, Draft Office Report, June 1.

Fisher, F.W. 1994. Past and present status of Central Valley Chinook salmon. Conservation Biology 8: 870-873.

Giorgi, A. E., T. W. Hillman, J. R. Stevenson, S. G. Hays, and C. M. Peven. 1997. Factors that influence the downstream migration rates of juvenile salmon and steelhead through the hydroelectric system in the mid-Columbia River basin. North American Journal of Fisheries Management 17:268-282.

Greene, S., California Department of Water Resources, Division of Environmental Services. Memo Report to R. L. Brown, Division Chief, DWR Division of Environmental Services. Re: Estimated winter-run chinook salmon salvage at the State Water Project and Central Valley Project Delta pumping facilities. May 8, 1992

Hallock, R. J. 1989. Upper Sacramento River steelhead, (Oncorhynchus mykiss), 1952-1998. Report to the Fish and Wildlife Service. 85pp.

Harvey, Brett. 2011. Length-at-Date Criteria to Classify Juvenile Chinook Salmon in the California Central Valley: Development and Implementation History. IEP Newsletter, Volume 24, Number 3, Summer 2011. 26-36

Healey, M.C. 1991 Life history of Chinook salmon (Oncorhynchus tshawytscha). In: Pacific salmon life history. Edited by: Groot, C. and Margolis, L. L. 311-394. Vancouver: University of British Columbia Press.

Kemp, P. S., Gessel, M. H., \& Williams, J. G. (2005). Fine-scale behavioralresponses of Pacific salmonid smolts as they encounter divergence andacceleration of flow.Transactions of the American Fisheries Society,134,390-398.

Kennen, J.G., S.J. Wisniewski, N.H. Ringler, and H.M. Hawkins. 1994. Application and modification of an auger trap to quantify emigrating fishes in Lake Ontario tributaries. North American Journal of Fisheries Management 14: 828-836.

Kimmerer, W. J. 2008. Losses of Sacramento River Chinook salmon and Delta smelt to entrainment in water diversions in the Sacramento-San Joaquin Delta. In: San Francisco Estuary and Watershed Science 6 (2).

Michel, C. J., A. J. Ammann, E. D. Chapman, P. T. Sandstrom, H. E. Fish, M. J. Thomas, G. P. Singer, S. T. Lindley, A. P. Klimley, and R. B. MacFarlane. 2013. The effects of environmental factors on the migratory movement patterns of Sacramento River yearling late-fall run Chinook salmon (Oncorhynchus tshawytscha). Environmental Biology of Fishes 96:257-271.

Mount, J., W. Bennett, J. Durand, W. Fleenor, E. Hanak, J. Lund, and P. Moyle. 2012. Aquatic ecosystem stressors in the Sacramento-San Joaquin Delta. San Francisco: Public Policy Institute of California.

Moyle, P. B. 2002. Inland Fishes of California. University of California Press, Canada. Pages 251263, 271-282.

National Marine Fisheries Service (NMFS), Southwest Region. 2009. Biological opinion and conference opinion on the long-term operations of the Central Valley Project and State Water Project.

Nobriga, M. and F. Feyrer. 2007. Shallow-water piscivore-prey dynamics in California's Sacramento-San Joaquin Delta. San Francisco Estuary and Watershed Science 5(2).

Orth, D. J. 1983. Aquatic Habitat Measurements. Pages 61-84 in: L. A. Nielsen and D. L. Johnson eds. Fisheries Techniques. American Fisheries Society, Bethesda, Maryland.

Palmer-Zwahlen, M., Gusman, V., Kormos, B. 2019. Recovery of Coded-Wire Tags from Chinook Salmon in California's Central Valley Escapement, Inland Harvest, and Ocean Harvest in 2015. Technical Report. December 2019

Volkhardt, G. C., S. L. Johnson, B. A. Miller, T. E. Nickelson and D. E. Seiler. Rotary Screw Traps and Inclined Plane Screen Traps. Pages 235-266 in D. H. Johnson, B. M. Shrier, J. S. O'Neal, J. A. Knutzen, X. Augerot, T. A. O'Neil, and T. N. Pearsons. 2007. Salmonid field protocols handbook: techniques for assessing status and trends in salmon and trout populations. American Fisheries Society, Bethesda, Maryland.

Yoshiyama, R. M., F.W. Fisher, and P.B. Moyle. 1998. Historical abundance and decline of Chinook salmon in the Central Valley region of California. North American Journal of Fisheries Management 18: 487-521.

Yoshiyama, R. M., E. R. Gerstung, F.W. Fisher, and P. B. Moyle. 2001. Historical and present distribution of Chinook salmon in the Central Valley drainage of California. Pages 71-76 in R. L. Brown, editor. Contributions to the Biology of Central Valley Salmonids. California Department of Fish and Wildlife, Fish Bulletin 179.

*CDEC gages: Sac. River below Wilkins Slough (WKL) for Knights Landing and Sac. River at Colusa (COL) for Tisdale Weir
**Sampling during high flows will be conducted depending on equipment, personnel safety, and logistical concerns. Sampling will be evaluate in real-time and may be discontinued for any of these reasons as well as if lethal take risk for listed species is high.
***High Flow period operations will be evaluated in real-time and may very with data needs, take risk, and equipment and personnel safety .
**** May have implications on trap capture efficiency and data comparability between sampling periods. Sampling will be conducted in a manner to allow for calculation of 24 hr. catch indices if possible.

Condition Dependent Sampling Schedule used to guide RST operations during varying environmental conditions.


[^0]:    ${ }^{1}$ Conducted by the California Department of Fish and Wildlife and funded by the Interagency Ecological Program

